



HAL
open science

Expanding networks of field hedges in densely populated landscapes in the Sahel

Pierre Hiernaux, Kalilou Adamou, Oumarou Moumouni, Matthew S Turner, Xiaoye Tong, Patrice Savadogo, Éric Mougin, Oumarou Malam Issa

► To cite this version:

Pierre Hiernaux, Kalilou Adamou, Oumarou Moumouni, Matthew S Turner, Xiaoye Tong, et al.. Expanding networks of field hedges in densely populated landscapes in the Sahel. *Forest Ecology and Management*, 2019, 440, pp.178-188. 10.1016/j.foreco.2019.03.016 . ird-02117954

HAL Id: ird-02117954

<https://ird.hal.science/ird-02117954v1>

Submitted on 8 Jul 2019

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1
2 **Expanding networks of field hedges in densely populated landscapes in the Sahel**

3
4 Pierre Hiernaux^{a*}, Kalilou Adamou^b, Oumarou Moumouni^b, Matthew D. Turner^c,
5 Xiaoye Tong^d, Patrice Savadogo^e, Eric Mougin^f, Oumarou Malam Issa^g.

6
7 a Pastoral Conseil, 30 chemin de Jouanal, 82160 Caylus, France ;

8 pierre.hiernaux2@orange.fr, * corresponding author

9 b Groupe de Recherche, d'Etudes et d'Action pour le Développement (GREAD), Niamey,
10 Niger ; adamou.kalilou@yahoo.fr,

11 c Department of Geography, University of Wisconsin, Madison, USA; mturner2@wisc.edu,

12 d Department of Geosciences and Natural Resource Management (IGN), University of
13 Copenhagen, 1350 Copenhagen, Denmark; xito@ign.ku.dk,

14 e World Agroforestry Centre (ICRAF), West and Central Africa Regional Office - Sahel
15 Node, BP E5118 Bamako, Mali; P.Savadogo@cgiar.org,

16 f Geosciences Environnement Toulouse (GET), Observatoire Midi-Pyrénées, UMR 5563
17 (CNRS/UPS/IRD/CNES), 14 Avenue Edouard Belin, 31400 Toulouse, France ;
18 Eric.MOUGIN@Get.omp.eu,

19 g Institut d'Ecologie et des Sciences de l'Environnement de Paris (IEES-Paris), UMR 242
20 (IRD, SU, CNRS, INRA, USPC, UPEC) 32 av., H. Varagnat, 93143 Bondy, France ;
21 oumarou.malamissa@ird.fr,

22

23

24

25 **Abstract**

26 Changing woody plant density across agricultural landscapes of the Sudano-Sahelian region is
27 a debated issue. This paper reports the results of an investigation on the contribution of field
28 hedges to overall woody plant density. Hedges separating individual cropped parcels were
29 studied within village agropastoral territories in the Dantiandou district in western Niger. In
30 1992, field hedges were mapped over the study area using aerial photography and in 2016,
31 using high resolution Google Earth imagery. In 1992, field hedge length was equal to 1006
32 km within 435 km² of croplands, equivalent to 2.3 km km⁻². In 2016, 17.5% of these hedges
33 had disappeared, but 1591 km of new hedges were observed resulting in an increased density
34 of hedges to 5.6 km km⁻². In 24 years, hedges had increased at a mean annual rate of 3.7%
35 likely associated with the splitting of crop fields by inheritance.

36 The composition and productivity of hedges were also monitored in 1996, 2010 and
37 2015. All trees (maximum height $\geq 4\text{m}$), shrubs ($< 4\text{m}$) and tussock perennial grasses were
38 recorded within twelve field hedge samples of 200m each. Field measurements were used to
39 estimate basal area, crown area, foliage mass and wood mass of each woody plant within the
40 sampled hedges. No significant trends were found between 1996 and 2015 in woody plant
41 density, basal area, crown cover, wood and foliage masses across all monitored hedges.
42 However, overall means hide a slight decrease in tree contributions, while shrub contributions
43 first increased and then decreased. They also mask contradictory trends among sites most
44 likely related to different rates of shrub coppicing and tree cutting. The woody species
45 composition of the hedges is poor with an increase of *Combretum glutinosum* to the detriment
46 of *Guiera senegalensis* over the study period.

47 Multiplying the hedge sample statistics by the changing lengths of field hedges in the
48 study area provides estimates of the contribution of the hedge woody plants to the woody
49 plant population at a landscape scale. Between 1992 and 2016, field hedges contributed to
50 increases of woody plant density by 3.9%, basal areas by 5.4% , crown cover by 2.7% ,
51 leading to 6.1% annual increase in foliage mass and 8.8% increase in wood mass.

52

53 **Key words:** agroforestry; hedgerow; land tenure; deforestation; bocage; tropical Africa

54 **Introduction**

55 The trend of woody plant populations in Sahelian West Africa is a topic of considerable
56 uncertainty and controversy. Many publications have referred to alarming deforestation
57 (Chamard and Courel 1999) caused by cropping or forestry exploitation to provide fuel to
58 urban centers (Peltier et al. 2009). However, a recent remote sensing work has found that
59 woody plant cover has generally increased across Sahel from 2000 to 2014 with a few
60 exceptions (Brandt et al. 2016). Within this debate, a major area of uncertainty is how woody
61 plant populations have changed on agricultural landscapes that represent a significant
62 proportion of the land area in the region.

63 The landscapes of the southern Sahel are strongly shaped by a long history of agrarian
64 management (Raynaut 2001; Ballouche and Rasse 2007). Human management reshapes rural
65 landscapes through land use (**Fig. SI 1**) with a major distinction being between rangelands
66 and agricultural lands, reviving the ancient distinction between saltus and ager (Poux et al
67 2009). Rangelands are defined by their pastoral management under communal access, but also
68 by their marginal potential for cropping associated with shallow or infertile soils, or with
69 increased flood risks in low lying areas. Moreover, rangelands cannot be distinguished from
70 agricultural lands simply by the presence of grazing since all portions of the rural landscape
71 are open to grazing after crop harvest (Turner et al. 2005). Gathering and wood collection also
72 extend to the whole landscape and are regulated by specific use rights (Sidikou 1997).
73 Agricultural lands include all the arable lands cropped or fallowed centered around villages.
74 Two components constitute the woody population in agricultural lands: 1. the large trees and
75 coppiced bushes of the agrarian parkland within fields and fallows (Boffa 1999), and 2. the
76 web of hedges that delineates fields, tracks and paths. Less spectacular and studied than the
77 trees (Gijsbers et al. 1994; Boffa 1995; Nouvellet et al. 2006; Yossi et al. 2006) and coppiced
78 bushes (Diack et al. 2000; Achard et al. 2001; Dick et al. 2010; Diakhaté et al. 2016) of the
79 agrarian parkland, the woody field hedges frame some of the most densely populated agro-
80 pastoral landscapes in West Africa (Pélissier 1953; Portères 1965; Seignobos 1980; Levasseur
81 2003).

82 The objective of this paper is to quantify changes over the past two decades in the
83 woody population of field hedges within a study area located within western Niger. Western
84 Niger is one of the areas of the Sahel where woody plant cover has been found not to have
85 increased from 2000 to 2014 (Brandt et al. 2016). Within this area, woody plant cover has

86 declined on plateau rangelands¹, declined slightly or remained the same on slopes, and
87 increased slightly in valleys and near villages. While slight positive trends have been
88 observed locally (San Emeterio et al. 2013), one would expect a decline of woody cover on
89 agricultural land due to the progressive expansion of the land cropped to the detriment of
90 fallows (Bouzou Moussa et al 2011; Tong et al 2017). Still, the coarse resolution of the
91 satellite imagery used in assessments of woody plant cover at the regional scale does not
92 allow a distinction of the contribution of field hedges from that of the parkland within field or
93 fallows.

94 In Western Niger, the trees, shrubs, and certain perennial grasses, which constitute the
95 field hedges, are not planted like in other regions of the Sahel (Yossi et al. 2006). Instead,
96 hedges are formed by selective sparing of woody plants at field edges when fields are cleared
97 (**Fig. SI 2**). The initial objective of this protection is to mark the field limits. The management
98 of hedges is conducted largely independently of the management within the surrounding
99 fields or fallows.

100 Landscape-level changes in the woody population of field hedges were estimated by
101 combining detailed ecological monitoring work of the composition and density of field
102 hedges with their mapping, using aerial photos and high resolution satellite imagery. Woody
103 populations of rangelands, fields and fallows in the region have most commonly been
104 assessed through either exhaustive census on sampled areas (Hiernaux et al. 2009) or by
105 distance methods (Picard and Bar-Hen 2007). However, the linear and dense arrangement of
106 the woody plants along field hedges is not compatible with either of these methods and
107 therefore, a more appropriate assessment method was developed for this research, as reported
108 below. Finally, observed changes in the spatial pattern of hedges will be related to changes in
109 land use and land tenure as documented by surveys of families farming fields in the study
110 area.

111

112 **Materials and method**

113 ***1.1. The study site***

¹ In western Niger, deforestation on the plateaus has often been observed within the patterned woody vegetation type known as ‘tiger bush’ or ‘spotted bush’ (d’Herbes et al. 1997). Degradation of tiger bush has been observed elsewhere in the Sahel (Abdourhamane Touré et al. 2010, Trichon et al. 2018).

114 To better understand the evolving ecology and contribution of field hedges to the overall
115 woody population in the Sahel, research was conducted within a 499 km² area in the
116 Dantiandou district of southwestern Niger (**Fig. 1**). In the study area, rangelands are confined
117 to the Oligocene ferricrete that caps the sedimentary plateaus at the highest elevations of the
118 landscape. Rangeland also extends along the edges of the plateau and on a few shallow soils
119 on the Pleistocene ferricrete downslopes (d'Herbès and Valentin 1997). Cropping is prevalent
120 on the sandy soils on the slopes and in the valleys, and on sandy clay soils lying in flats and
121 stream beds (Turner and Hiernaux 2015). Agricultural lands are composed of a minority of
122 permanently cropped fields, close to villages and to pastoral camps, and a majority of shifting
123 crop fields and fallows further away (Osbahr 2001). Indeed, permanent cropping requires
124 fertilization that is most commonly achieved by corralling livestock on the fields during the
125 dry season (Hiernaux and Diawara 2014) with or without the application of domestic wastes
126 carried to the field by carts. The capacity to manure depends on the size of the family herd
127 and on the extent of seasonal transhumance of the herd away from the village. The soil
128 fertility of shifting fields is managed by alternating cropping for five years with fallowing that
129 lasts typically three years but may last much longer especially on lands marginal for cropping
130 (Hiernaux and Turner 2002). Fallow duration also depends on the field's tenure status (Turner
131 and Moumouni 2018a). Agricultural lands are clustered around villages the locations of which
132 are where lands were free to crop and ground water was accessible to hand-dug wells (more
133 recently, boreholes) at the time of settlement (Favreau et al. 2002).

134 The study area's population is dominated by the Djerma who also control agricultural
135 land through village chieftaincies. Although Djerma families hold no formal land title, they
136 have legal customary rights to their cropped field; rights which are generally inheritable by
137 sons (Lavigne-Delville 2003; Turner and Moumouni 2018a). The founding Djerma families
138 of the village enjoy chieftaincy rights on lands often arranged in sectors radial from the
139 village. Among them, the village chief manages usufruct rights to the chieftaincy common
140 lands. These usufruct rights are assignable by contract to new settlers, including agropastoral
141 Fulani families who have settled in the vicinity of the villages within the study area dating
142 back to the severe 1973-74 drought. These rights are restricted to cropping activity without
143 exclusive rights to exploit woody plants or to graze livestock on the agricultural land.

144 The landscape is further structured by a web of tracks that radiate from each village to
145 access the fields and travel to neighboring villages. In addition, there is a web of livestock
146 paths that allow local livestock access to water points and rangelands, especially during the
147 wet season when they are not authorized to cross cropped fields. These livestock paths also

148 connect to a web of regional livestock paths allowing regional and international transhumance
 149 (Turner et al. 2014). The pastoral camps of settled Fulani families are positioned along these
 150 livestock paths just outside the edge of permanently cropped fields surrounding each village.
 151 Their positions reflect that cropped fields are off limits to livestock during the growing season
 152 and that pastoral families depend on village water resources during the dry season.

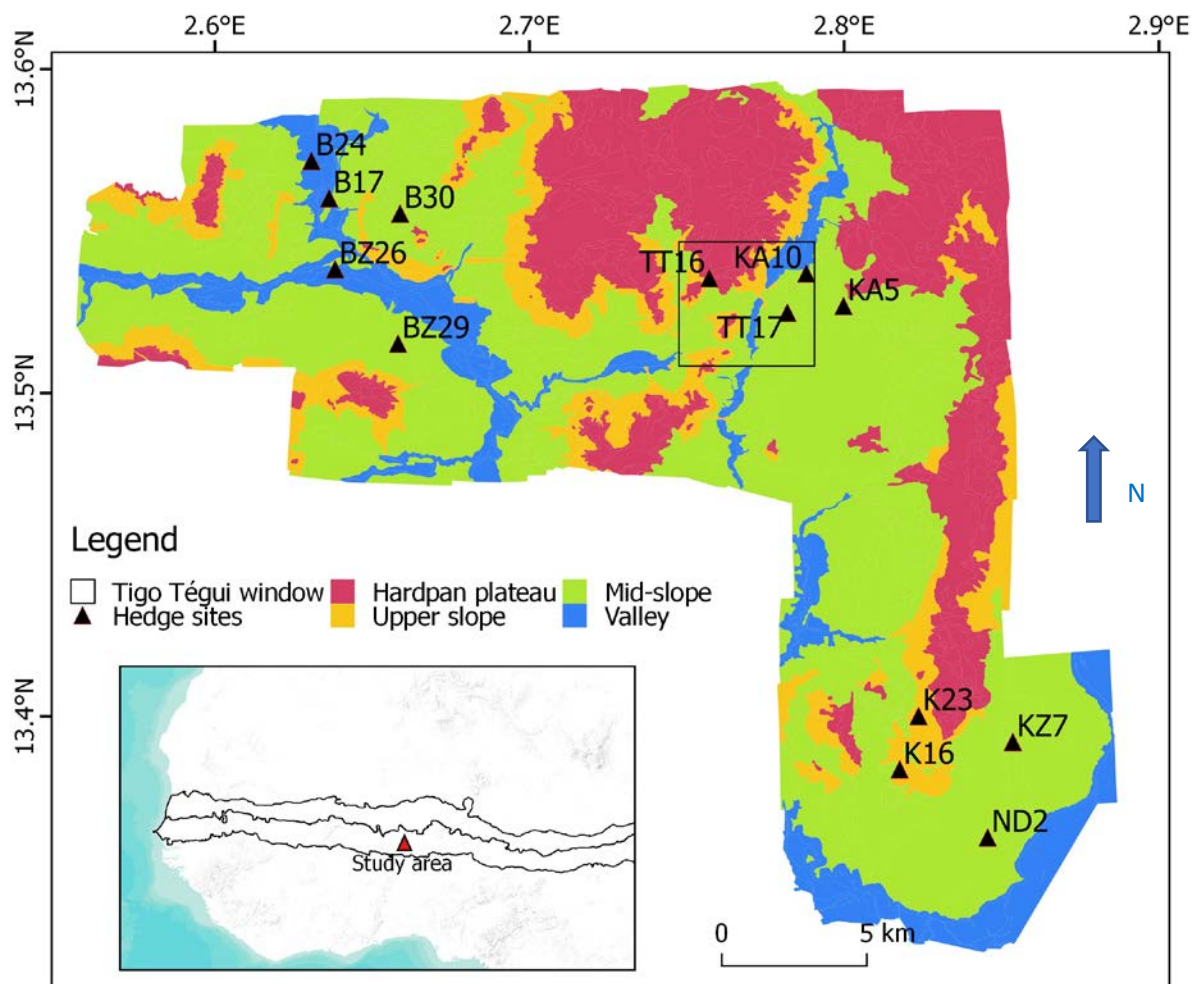


Figure 1 Geographic location of the study area (499km²) in the Dantiadou District of Niger. The three lines of the small maps figure isohyets 150, 300 and 600mm from CHIRPS data 1982-2014 delineating northern, central and southern Sahel bioclimates. Twelve field hedges monitored since 1996 (+ B17 only observed in 1996) are located relative to the main edaphic situations: hard pan plateau, top sandy slope, mid and bottom sandy slopes and valleys (Hiernaux and Ayantunde 2004). The black square frame locates the more detailed map of **Fig 6**.

153 The study area's climate is typical for the semi-arid monsoonal tropics with an average
 154 annual rainfall of 494 ± 30 mm (1990-2018) at Banizoumbou, a village within the study area,

155 distributed from June to October in 35 ± 6 rainy days most of which as convective storms
 156 (Panthou et al. 2014). There are large rainfall variations between years with series of wetter
 157 and dryer years such as the regional droughts of the early 1970's, mid 1980's and more
 158 recently between 2009 and 2011 (Lebel and Ali 2009). However, there is no overall trend in
 159 annual rainfall over the 1990-2018 period that encompass the monitoring period for this study
 160 (Fig. 2).

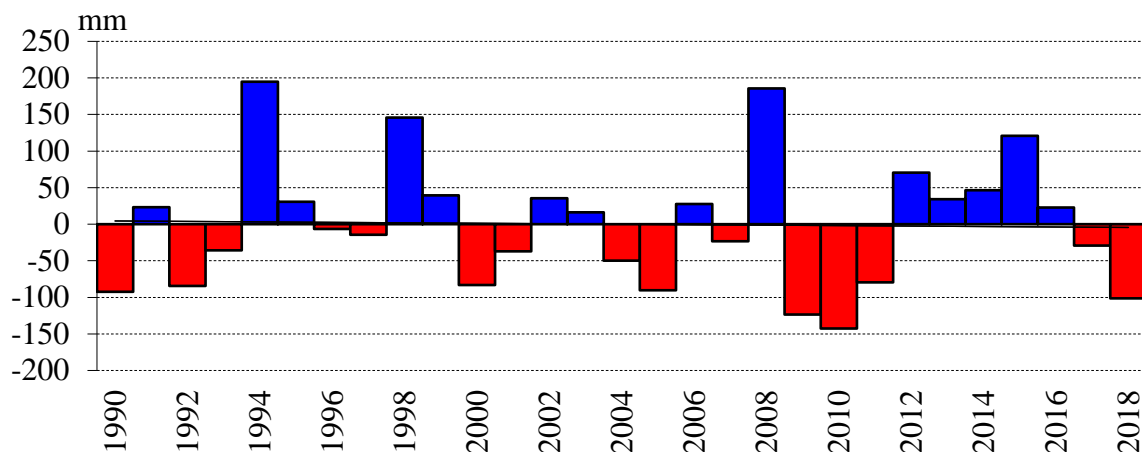


Figure 2 Annual rainfall departure from the mean over 1990 to 2018 (494 ± 30 mm) at Banizoumbou ($13^{\circ}32'N$, $2^{\circ}40'E$).

161 **1.2. Mapping land use and field hedges**

162 Woody field hedges were initially mapped by stereoscopic photointerpretation of 1992
 163 panchromatic aerial photos (**Table SI 1**) over the study area. The map was digitized in
 164 ATLAS-GIS and geometrically corrected using 150 GPS readings (Trimble Pathfinder Basic
 165 Plus GPS receiver) under a UTM projection with a WGS-84 datum. Yet, when the file was
 166 overlaid on Google Earth small shifts in random directions were observed and these were
 167 corrected manually. The woody field hedges in 1992 were mapped along with six classes of
 168 land use (built-up area, manured cropland, unmanured cropland, recent fallow, old fallow,
 169 rangeland) together with tracks, cattle paths, pastoral camps, granaries, isolated trees,
 170 ephemeral water courses, ponds, wells and boreholes (Hiernaux and Ayantunde 2004). The
 171 woody field hedges in 2013-16 were mapped using available Google Earth imagery (**Table SI**
 172 **1**). The hedges mapped in 1992 were first overlaid on the google map imagery and the hedges
 173 that had disappeared were identified and erased from the 2016 file. Then the hedges not
 174 mapped in 1992 but present in 2013-2016 were mapped in a separate file. To document the
 175 land use changes, the categories of land use mapped in 1992 plus an additional category of

176 “pristine savanna” (no sign of cropping at least during the previous century) were mapped
177 over the same area covered by historical aerial photo coverages of 1950 (IGN) and 1975
178 (IGNN). Land use mapping was also performed on low altitude aerial color slides taken in
179 1994, 1995 and 1996. Finally, land use mapping was extended to high resolution multispectral
180 satellite data (SPOT images) by supervised classification in 2007, 2008, 2009, and 2010.

181

182 1.3. *Monitoring woody plant and tussock grasses in field hedge samples*

183 Twelve samples of field hedges were selected in 1996: three on plateau dunes or on sandy
184 upper slopes; three on sandy mid-slopes; three on loamy-sand bottom-slopes; and three on
185 alluvial soils in valleys (**Fig. 1**). Only one of the sites (K23) is located along permanently
186 cropped fields. All the others are along fields that are managed in a crop-fallow cycle. First
187 measured in 1996, the hedges were revisited in 2010 and 2015. One of the hedges described
188 in 1996 had disappeared in 2010 due to land erosion on a stream bank (B17) and was replaced
189 by a different hedge (BZ26).

190 The monitored field hedge sections were 200m long, subdivided into eight segments
191 of 25m. The perennial grasses (*Andropogon gayanus*, *Aristida sieberiana*), shrubs (maximum
192 crown height below 4m) and trees (crown height above 4m) along the hedges are
193 systematically measured. For each woody plant, the following were recorded: the species,
194 crown height, largest crown diameter (D, m), crown diameter perpendicular to the largest
195 diameter (d, m), and the circumference (C, cm) of each stem at 10 cm above ground level.

196 The density of shrubs and trees was calculated by simply counting individual plants
197 along the hedge section. Errors are very unlikely, limited to the possibility of inadequately
198 distinguish individual shrubs when highly intertwined. The basal area (Ba, cm²) of each
199 woody plant was calculated by summing the basal area of each stem calculated as :

$$200 \quad Ba = C^2 / (4 \cdot \pi) \quad (1)$$

201 Risks of error in assessing basal areas are minor as well, limited to missing small stems,
202 especially after recent shrub coppicing, and to imprecise measurements of the stem
203 circumferences, especially for low branching trees or shrubs.

204 The foliage dry mass (Fm) and wood dry mass (Wm) are estimated using existing
205 species-specific allometric functions of the circumference of each stem (Cissé 1980; Henry et
206 al. 2011):

207
$$F_m = \alpha \cdot C^\beta \quad (2)$$

208
$$W_m = \gamma \cdot C^\delta \quad (3)$$

209 The species-specific coefficients α , β , γ and δ are provided in **Table SI 2** . Error associated
210 with the use of allometric functions such as these is likely if the sizes of woody plants
211 sampled fall outside the range of circumferences used to generate the function. Unfortunately,
212 such ranges are not systematically documented for the allometric functions used but it is
213 unlikely that the size distribution of woody plants sampled here are outside the ranges used to
214 develop these functions.

215 The crown area (C_a , m^2) is estimated as the area of the ellipsoid shape defined by the
216 largest crown diameter (D , m) and the perpendicular diameter (d , m):

217
$$C_a = \pi \cdot (D \cdot d) / 4 \quad (4)$$

218 There is a possible observer bias tied to the crown diameter estimation being either measured
219 in 1996 and 2015, as the extreme longest diameters, between the more remote branchlets,
220 while in in 2010, the diameter estimated from a rounded ellipsoid fitting the crown.

221 Densities, basal areas, crown areas, wood and foliage masses are expressed per 100m of
222 hedge. The contribution of field hedge woody plants to the overall woody plant population is
223 assessed in two different ways. The first approach estimates the contribution per unit area (ha)
224 by relating the measurements expressed by 100m of hedge to the 10 m-wide band surrounding
225 the hedge, i.e. 5 meters on either side of the hedge axis. This is justified by the observation
226 that the maximum crown diameter of the trees in monitored hedges did not exceed 10 meters.
227 Using this approach, leaf and wood masses per hectare equivalent are estimated by simply
228 multiplying by 10 the values given for 100 meters of hedge. The second way is to estimate the
229 absolute number of woody plant individuals contributed by hedges by multiplying the hedge
230 woody plants values expressed by 100 meter of hedge by the length of field hedges mapped
231 on the same date.

232 **1.4. Surveying land tenure and human population**

233 The land-use maps in 1994 to 1996 were used as background to survey and map land tenure,
234 identifying the family managing each field and under which tenure status: primary rights to
235 the land generally held by members of chieftaincy lineage and the few purchases of land, or
236 secondary rights which characterizes the rights held by those outside of the chieftaincy
237 lineage (Hiernaux and Ayantunde 2004). An additional survey was conducted in 1998 on the

238 history of each field, establishing the ancestor of the actual family head that had first cleared
 239 the field for cropping, the year of that clearing and the list of the names of the successive
 240 managers since clearing. The land tenure status of crop fields and land rights of families in
 241 three major villages have been further studied and characterized recently (Turner and
 242 Moumouni 2018a, b). The composition of families were surveyed in 1994, 1998 and 2013
 243 (Zezza et al. 2016) that allows an assessment of demographic growth and changes in family
 244 composition. These population surveys were compared with the more recent administrative
 245 population census performed in 2008 and 2013 (INS-Niger 2013).

246

247 2. Results

248 2.1. Field hedge and land use mapping

249 *Field hedge mapping*

250 The total length of woody field hedges in the study area in 1992 equals 1006 km over 499
 251 km² of land (equivalent to 2.02 km km⁻²) and over 435 km² of agricultural land (2.31 km km⁻²
 252) when the unarable rangelands, without field hedges, are excluded (**Fig. 3, Table 1**). In 2016,
 253 174.1km (17.5%) of these 1992 field hedges had disappeared, but 1591.3 km of new hedges
 254 were observed so that the resulting density of woody field hedges per area of agricultural land
 255 reached 5.56 km km⁻² (**Fig. 4**). Woody field hedges had thus increased by a factor 2.4
 256 equivalent to a mean annual increase rate of 3.7% from 1992 to 2016.

257 **Table 1.** Inventory of woody field hedges over the study area (499 km² out of which 435 km²
 258 of agricultural lands) in 1992 and 2013-2016. The density of hedges is calculated at the full
 259 landscape (499 km²) or at the agricultural land (435 km²) scale.

Year	Type of hedges	Number	Total length	Mean length	Density within study area	Density in study area's agricultural land
		#	Km	m	km/km ²	km/km ²
1992	Hedges	1981	1005.7	507	2.02	2.31
2013-16	1992 Persistent	1716	829.6	483	1.66	1.91
2013-16	New hedges	4682	1591.3	340	3.19	3.66
2013-16	All hedges	6398	2420.9	378	4.85	5.56

260

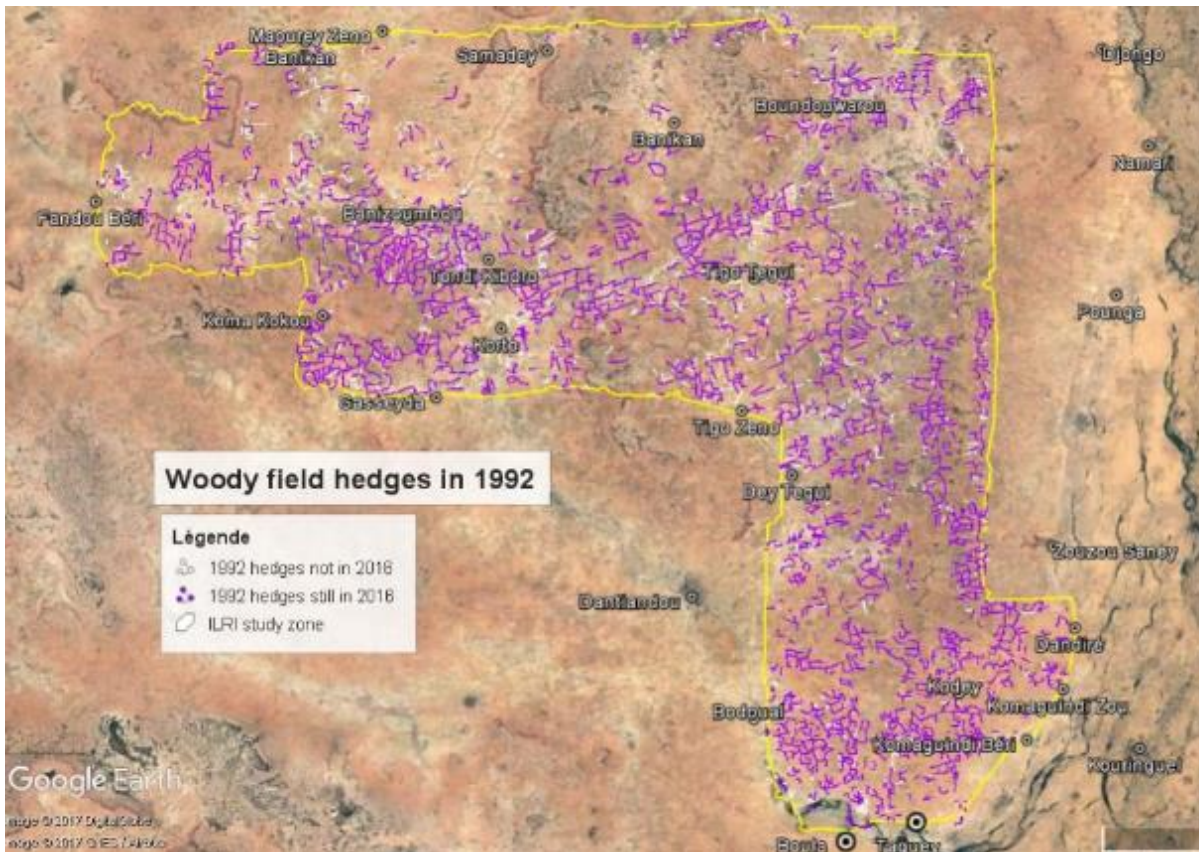


Figure 3. Map of the woody field hedges in 1992 overlaid on Google Earth image. Hedges that were eliminated (white) and maintained (purple) by 2013-16 are delineated with the study area delineated in yellow.

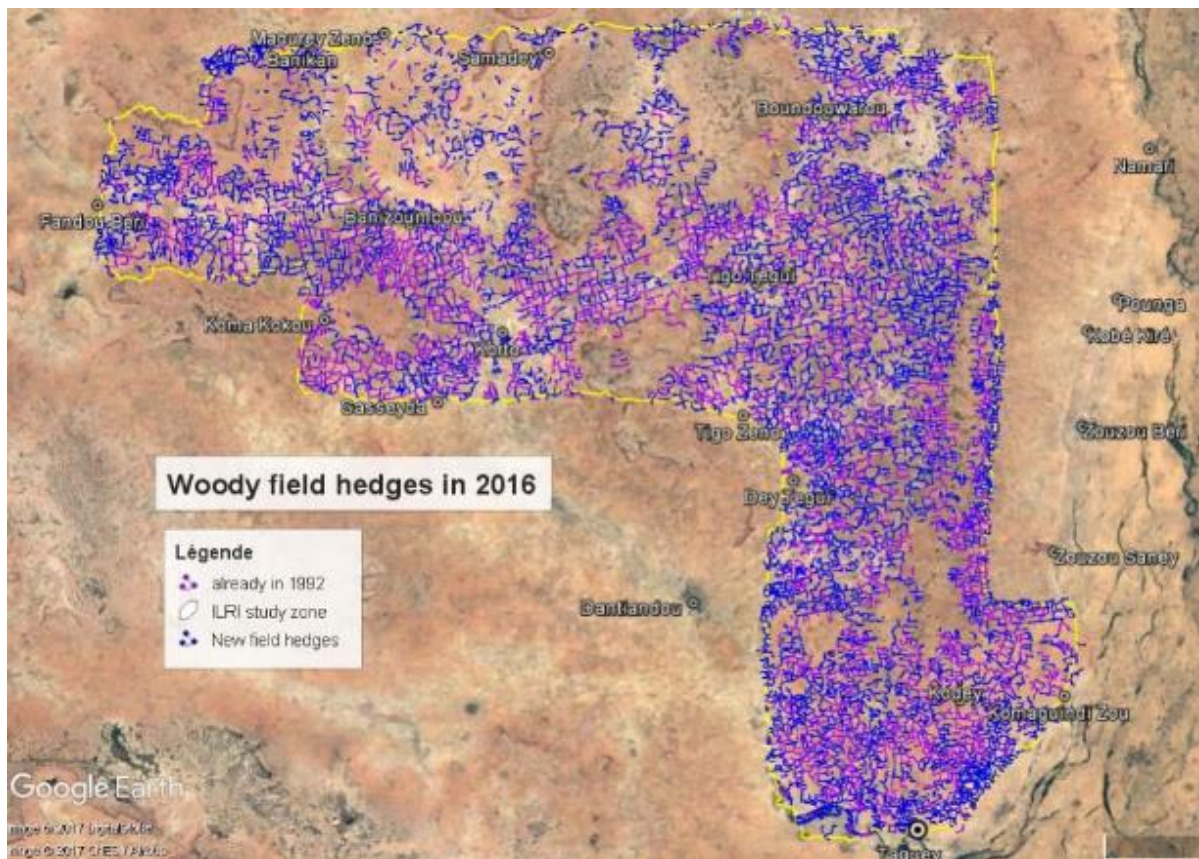


Figure 4. Map of the woody field hedges in 2013-16 from Google earth imagery of 11/11/2016 (70% East side), 31/01/2016 (10%, Mid-Center), or 28/09/2013 (20%, West). Hedges that were already present in 1992 (purple), new hedges (blue) are delineated with study area delineated in in yellow.

261

262 ***Land use mapping.***

263 There were significant changes in land use within the 499km² study area since 1950. There
 264 has been a rapid extension of the cropped area from 1950-1975 with the disappearance of
 265 ‘pristine’ savannas (**Fig. 5**). Since 1992, cropped area continued to increase to the detriment
 266 of fallows with interannual variations due to farmers adjusting their sowing and weeding
 267 practices to observed rainfall received by individual fields (**Fig. 2**). The land use dynamics
 268 from 1950 to 2018 is also illustrated by series of aerial views centered over the northern
 269 portion of one village’s fields located in the center of the study area (**Fig SI 2**).

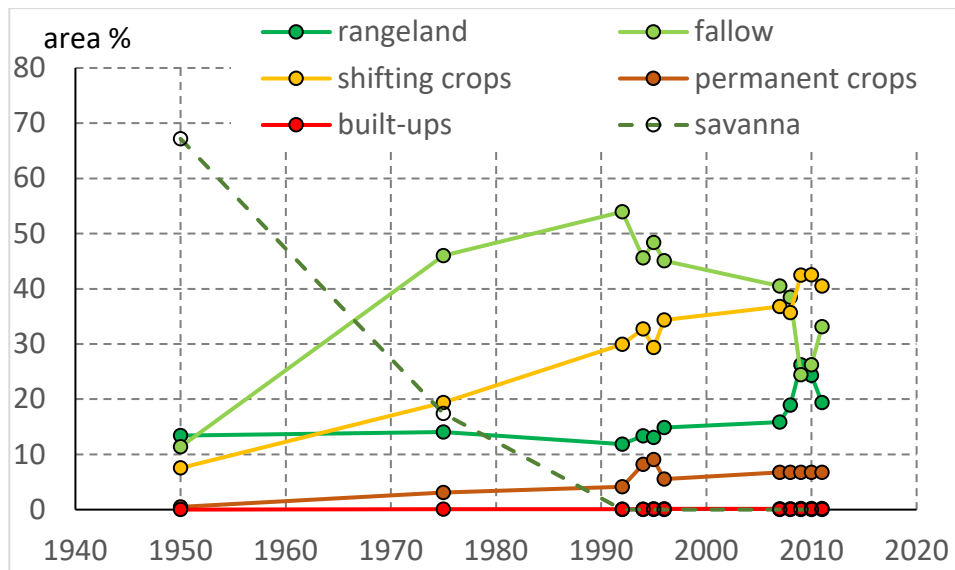
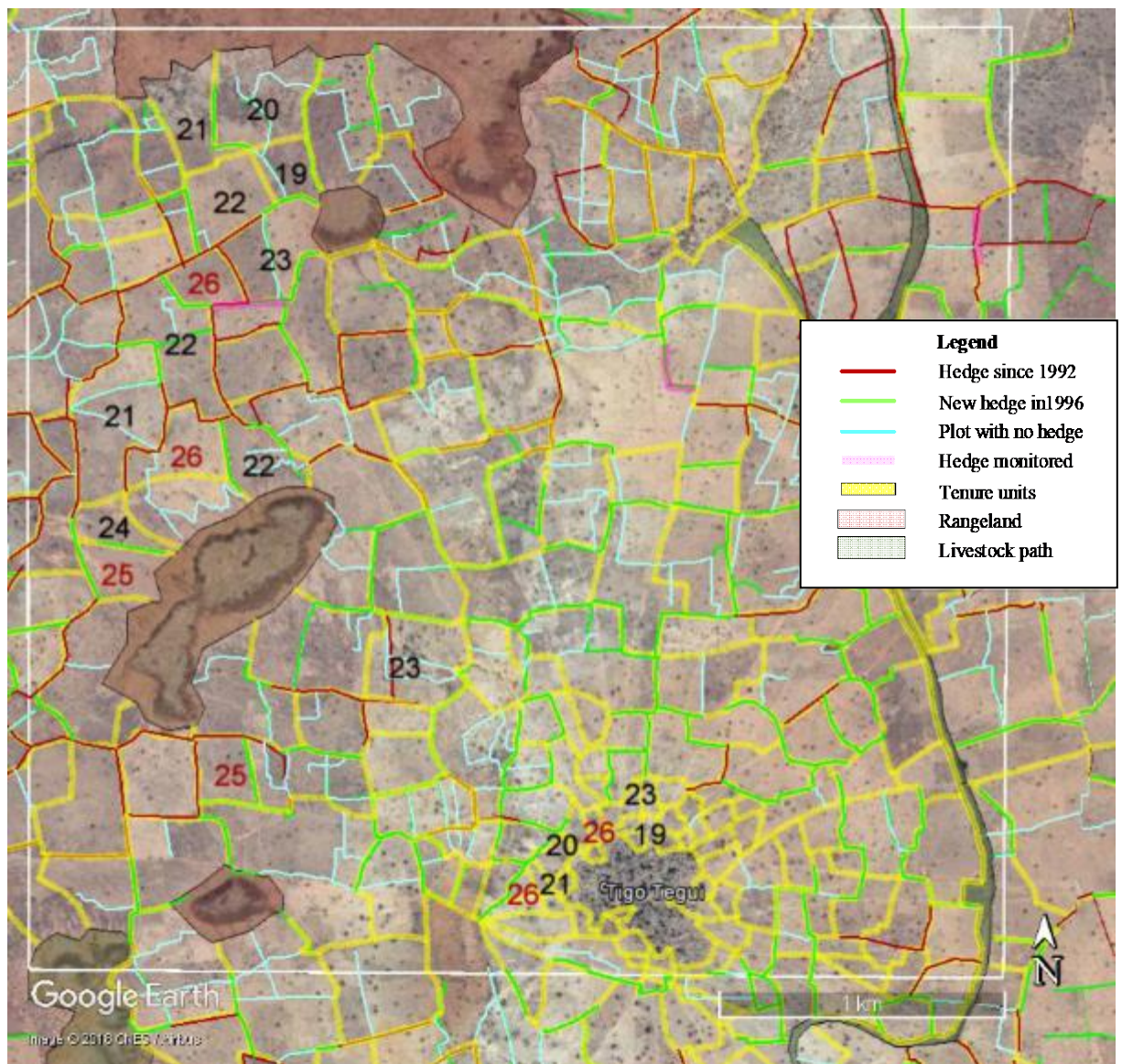


Figure 5. Changes in land use over the study area (499km²). Areas are estimated from classification of aerial photos (1950-1996) and supervised classifications of SPOT imagery (2007 onwards). The cropland is categorized as either shifting crop fields or permanently cropped fields.

270 *Land tenure mapping and recording field management history*

271 Using the 1994 land-use map, interviews of farmers produced a land tenure map as described
 272 above. **Figure 6** shows a portion of the land tenure map, overlaid on the map of the woody
 273 field hedges for an area lying northwest of a village in the center of the study area (see inset of
 274 **Figure 1** for this figure's location within study area). The process of agricultural land splitting
 275 through intergenerational inheritance is illustrated by a few fields identified on this portion of
 276 the land tenure map (with fields attached to certain families designated by family codes 19 to
 277 26 described more fully in **Fig SI 9**). In 1998 these fields were separately managed by eight
 278 families (seven brothers and one of their sons) -- all descendants from a common ancestor,
 279 Boulymane K., grandson of Koutoubi H. who founded the village in 1906 (Turner and
 280 Moumouni 2018a). Boulymane K.'s only son, Moussa B., had seven sons who shared his
 281 rights to the land he inherited (**Fig. SI 9**). The next generation currently in charge of field
 282 management is composed of 26 family heads, associated with further splits of the land parcels
 283 by a factor 3.7 on average. Some of these plot partitions generate new field hedges, either
 284 already established in 2016 (in light green in **Fig. 6**) or just outlined (in light blue). Indeed,
 285 the split of fields between brothers is sometimes initiated before the father's death (8.1 ± 4.7
 286 years before, on average) but more commonly, after the father's death (13.4 ± 16.6 years after,
 287 on average) (Turner and Moumouni, 2018b).



289

Figure 6. Land tenure (yellow thick polygons as recorded in 2002), woody field hedges (since 1992 in dark red and new in 2016 in green), and recent field partition with no hedge yet (light blue), to the northwest of a village within the study area. The numbers refer to related farmer families (**Fig. SI 9**). The map is overlaid on high resolution Google Earth satellite image dated 11/11/2016.

290

291 **2.2. Field hedge monitoring**

292 ***Woody plant density, basal area, crown cover and foliage masses in the field hedges.***

293 Mean woody plant density of surveyed hedges varies little between 15.9 in 1996, 20.7 in 2010
 294 and 16.6 per 100m of hedge in 2015 (**Table 2; Fig. SI 4a**). In other words, woody plants, on
 295 average, are standing 5 to 6 meters apart along the hedge (**Fig. SI 1**).

Table 2. Mean density, basal area, crown cover, foliage and wood masses per 100m of field hedge measured in 12 sites in 1996, 2010 and 2015.

Parameter for 100m of field hedge	Year	Trees		Shrubs		All woody	
		mean	Stdev	Mean	Stdev	mean	Stdev
Density (#)	1996	4.0	1.9	11.8	8.2	15.8	6.8
	2010	2.5	1.8	18.1	9.1	20.5	8.2
	2015	2.2	2.0	14.3	6.8	16.6	6.4
Basal area (m ²)	1996	0.2	0.1	0.3	0.2	0.2	0.1
	2010	0.2	0.2	0.2	0.1	0.4	0.2
	2015	0.2	0.2	0.2	0.1	0.4	0.2
Cover (m ²)	1996	167.9	115.5	73.3	55.0	234.1	101.9
	2010	52.3	47.4	60.8	38.4	110.5	37.0
	2015	77.2	77.7	112.0	84.0	185.1	85.3
Leaf mass (kg dry mass)	1996	49.6	33.5	16.9	14.1	66.6	29.3
	2010	59.4	78.5	61.2	38.2	120.6	48.7
	2015	57.8	76.7	52.5	41.7	110.3	50.1
Wood mass (kg dry mass)	1996	678.1	510.2	148.0	127.0	826.1	439.8
	2010	1576.9	2177.5	581.0	409.5	2157.9	1941.6
	2015	1821.2	2637.6	755.3	1004.0	2576.5	2318.3

296
 297 **Table 2** shows that across the years 1996, 2010, 2015, basal area and leaf mass of all woody
 298 plants first increased and then decreased as plant density while wood mass increases
 299 throughout the monitoring period. The opposite trend is seen for cumulative crown cover
 300 which first declines and then partially recovers over the same time period . This observation
 301 likely results from the bias in crown diameter measurements in 2010 described above and
 302 should be thus disregarded.

303 Among sites, contrasting trends were observed throughout the period. Woody plant
 304 density decreased at two sites, remained relatively stable at four sites and increased from 1996
 305 to 2010 and then slightly decreased in 2015 at the other four sites (**Fig. SI 5a**). Similarly, the
 306 mean decrease in basal area hides an increase at five sites and a decrease at the other seven
 307 (**Fig. SI 5c**). Cumulative crown cover, leaf mass, and wood mass display much less inter-site
 308 variation with only one site in each case showing contrasting trends (**Figures S1 5d and S1**
 309 **5e**).

310

311 ***Tree and shrub contributions***

312 Crown cover, basal area, leaf mass, and wood mass show the same general trends for trees
313 and shrubs across the monitored period. However, different trends in plant density were
314 observed. The slight increase in woody plant density hides a decrease in the mean density of
315 trees while shrub density strongly increased between 1996 and 2010, and then declined (**Fig**
316 **SI 4a**). However, looking at each hedge, the overall decrease in tree density does not apply to
317 sites K23, TT16 and BZ29 where tree density and aggregated basal area increased between
318 2010 and 2015 probably because some of the shrubs passed the threshold of 4m in height and
319 were counted as trees (**Fig SI 6 a,c**). For the other hedges, the decrease in tree density, basal
320 area, with associated decreases in tree wood and leaf masses could be due to selective
321 logging, especially at B24, K16, KZ7 and ND2 (**Fig 6 a,c 7a,c**). Shrub coppicing, such as that
322 experienced at hedges B24, TT16 and TT17 between 2010 and 2015 did not affect shrub
323 density but reduced basal areas, wood and leaf masses. In contrast, the KA5 and ND2 hedges
324 that were not coppiced during the same period, show opposite trends (**Fig 6 b,d 7b,d**).

325 ***Botanical composition of field hedge woody populations.***

326 The number of woody plant species recorded in the 12 monitored hedges (**Table SI 3**) varied
327 between 9 in 1996, 14 in 2010 and 12 in 2015 among which, three species largely dominate,
328 together representing 93-98% of the total value depending on the parameter considered (**Fig.**
329 **8**): *Guiera senegalensis*, *Combretum glutinosum* and *Piliostigma reticulata*. As a shrub, *G.*
330 *senegalensis* ranks first in density and crown cover. *C. glutinosum*, which may grow as a tree,
331 ranks first in wood mass. *P. reticulata*, a small tree which is often coppiced, always ranks
332 third. Although plant density did not change much over time, *C. glutinosum* increased in
333 crown cover and basal area leading to increases in leaf and wood masses. On the contrary *G.*
334 *senegalensis* decreased in both crown cover and basal area leading to decreases in leaf and
335 wood masses from 1996 to 2015.

336

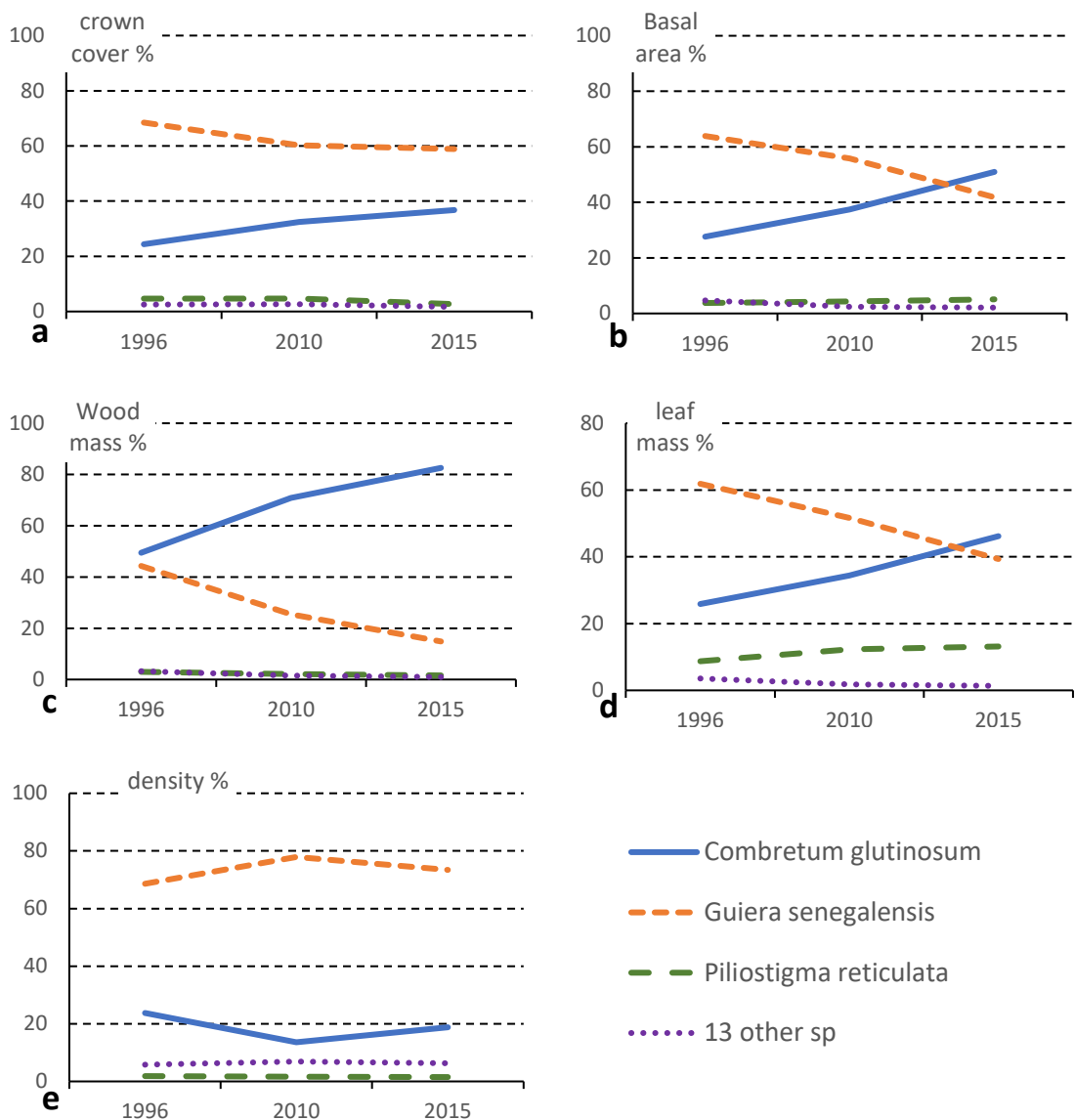


Figure 8. Species contribution (%) to the mean **a)** crown cover , **b)** total basal area, **c)** wood mass, **d)** leaf mass and **e)** woody plant density along 12 field hedges in 1996, 2010 and 2015.

337

338 *Two perennial tussock grasses in the field hedges*

339 Two perennial grass species are common in some of the field hedges, growing between
 340 woody plants. *Andropogon gayanus*, a 2-3m tall tussock grass, is protected by farmers in
 341 fields and particularly in the hedges because of its multiple uses (woven mats, thatch, chicken
 342 coops, beehives...) and the good fodder value of its dry season regrowth. *Aristida sieberiana*
 343 is not protected but its seed dispersion and germination is favored by soil disturbance
 344 associated with cropping. Both species' densities decreased during the study period with a
 345 faster reduction of *A. sieberiana* than *A. gayanus* (**Fig. 9**). The large standard deviation of

346 density reflects quite different values among sites, the perennial grasses being more frequent
347 either in more silty sands at the top of slopes or in depressions (**Fig. SI 8**). However a
348 decrease in perennial grasses occurred in all sites.

349

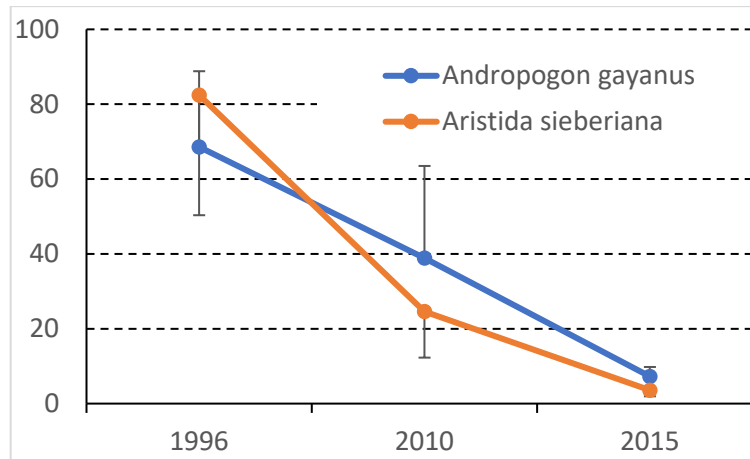


Figure 9. Mean and standard deviations of density (#/100 m of field hedge) of *Andropogon gayanus* and *Aristida sieberiana* tussocks along 12 field hedges in 1996, 2010 and 2015.

350

351 *2.3. Contribution of field hedges to woody plant populations at the landscape scale*

352 Relating the woody plants described for 200m of field hedge to a band of land of 10m wide
353 provides an estimate of the field hedge contribution to woody plant density, basal area, cover,
354 leaf and wood masses per unit land. These contributions expressed per hectare are calculated
355 by multiplying by 10 from the statistics expressed per 100 m of hedge (**Table 2**). Expressing
356 hedge contributions in this way allows one to compare these values with those estimated
357 within cropped or fallowed fields of the study area. The woody plant populations of the field
358 hedges are much denser than that of cropped and fallowed fields (San Emeterio Cabañes et al.
359 2013; Brandt et al. 2016), with densities and cover closer to that of Sahelian open woodlands
360 (Hiernaux et al. 2009).

361 Alternatively, multiplying the hedge statistics from field observations in 1995 (**Table**
362 **2**) by the density of hedges mapped in 1992 (**Table 1**) provides estimates of contribution of
363 hedges over the study area's agricultural land in 1992-95 (**Table 3**). Similarly 2015 field
364 statistics and 2013-2016 hedge map allows one to estimate the woody hedge contributions in
365 2013-16. Estimates of the parameter contributions of woody plants found within hedges
366 across the study area all increased over the 24 years with mean annual growth rates varying

367 between 3.9% for density, 5.4% for basal area, 2,7% for crown cover, resulting in annual
 368 increases by 6.1% in foliage mass and 8.8% wood mass. Shrubs are driving the increases in
 369 density, crown cover and foliage mass, while trees dominate the increase of basal area and
 370 wood mass.

371 **Table 3.** Mean and standard deviation of the contribution of the field hedges to the density (ha⁻¹)
 372 ¹), basal area (m² ha⁻¹), cover (%)), foliage and wood masses (kg ha⁻¹) of woody plant
 373 population on agricultural land within the study area in 1992-2005 and 2013-2016.

Parameter	Interval	Trees		Shrubs		All Woody	
		mean	Stdev	mean	Stdev	mean	stdev
Density ha ⁻¹	1992-95	0.92	0.44	2.73	1.89	3.65	1.57
	2013-16	1.22	1.11	7.95	3.78	9.22	3.55
Basal area m ² ha ⁻¹	1992-95	0.04	0.03	0.06	0.04	0.06	0.03
	2013-16	0.1	0.09	0.11	0.06	0.21	0.09
Cover %	1992-95	0.39	0.27	0.17	0.13	0.54	0.24
	2013-16	0.43	0.43	0.62	0.47	1.03	0.47
Foliage mass kg ha ⁻¹	1992-95	11.5	7.1	3.9	3.3	15.4	6.8
	2013-16	32.1	42.6	29.2	23.2	61.3	27.9
Wood mass kg ha ⁻¹	1992-95	156.6	117.9	34.2	29.3	190.8	101.6
	2013-16	1012.6	1466.5	419.9	558.2	1432.5	1289

374

375

376 3. Discussion

377 Literature on field hedges is scarce in the Sahel. Their existence in the agrarian landscape is
 378 recognized (Pélissier 1953; Sanogo et al 2000; Levasseur 2003) but there have been no
 379 quantitative descriptions of densities, cover or wood masses. The literature has focussed more
 380 on the impact of woody hedges on biodiversity or agroecosystem processes (Burel 1996),
 381 with a particular attention to their potential to control either wind (Banzhaf et al. 1992;
 382 Michels et al. 1998) or run-off erosion (Valet 2000; Smolikowski et al. 2001). The role of
 383 hedges serving as land tenure markers (Pélissier 1995) is also mentioned. Moreover, there is
 384 an abundant technical literature on the agroforestry methods, best adapted woody species,
 385 planting or drafting practices to establish new live hedges, as if they were not part of the
 386 agrarian system already (Depommier 1991; Ayuk 1997; Yossi et al. 2006). Reflective of this,
 387 the woody plants growing in the hedges are generally excluded from the woody plant surveys
 388 carried out in forests and savannas, and also in cropped or fallowed fields. One reason for this

389 disregard is the inability of standard survey methods to assess linear vegetation patterns. Yet,
390 as demonstrated by this research, woody plants found within field hedges contribute
391 significantly to the woody populations of agricultural landscapes.

392 **3.1.Heterogenous structures and homogeneous species composition**

393 Woody plant populations were found to differ significantly among the 12 sampled field
394 hedges with coefficients of variation of the 12 hedge means between 40 and 43% for woody
395 plant density, basal area, crown cover, foliage and wood masses (**Table 2**). The heterogeneity
396 increases when shrubs and trees are separated with coefficients of variation for shrubs ranging
397 between 55% for density and 71% for crown cover, and for trees between 71% for density
398 and 87% for crown cover. The larger heterogeneity of crown cover could be related to the
399 variations caused by the irregular coppicing of hedge shrubs. Because of their higher density,
400 shrubs contribute about as much as trees to total basal area, crown cover and foliage mass.
401 These variations are not explained by the geomorphic position nor related run-off/run-on
402 balances (not shown).

403 The species composition of the field hedges of the study area is relatively poor by
404 comparison with other woody populations in Niger (Mahamane et al. 2009; Idrissa et al.
405 2017) with only 16 species registered across the 12 sites over the three years of observation
406 (**Table SI 3**). This poverty is reinforced by the fact that three species: *G. senegalensis*, *C.*
407 *glutinosum* and *P. reticulata* together represent 93-94% of plant density, 96-97% of basal
408 area, 97-98% of crown cover, wood and foliage masses. In Western Niger, the trees and
409 shrubs of these field hedges are not planted, nor sown, but result from selective sparing of
410 woody plants at field clearing. They are just the most common species found on the infertile
411 acidic sandy soils in the southern Sahel (Hiernaux and Le Houérou, 2006)

412

413 **3.2. Weak trends in hedge woody plant and perennial grasses densities, cover and** 414 **production over last two decades.**

415 There is no significant consistent trend between 1996 and 2015 in hedge woody plant
416 density, basal area, crown cover, wood and foliage masses across the 12 hedges monitored.
417 Moreover, the weak trends found for total woody plant population in the hedges result from
418 different trends between trees and shrubs (**Fig. SI 4, 6, 7**) and between hedges (**Fig. SI 6, 7,**
419 **8**). The weakness of the overall trends, the contradictory trends among sites; between trees
420 and shrubs, and between the two main species all argue in favor of a woody population
421 dynamics more linked to site-specific management by selective shrub coppicing and tree

422 cutting than to long term climate changes. Given that the prime objective of the woody hedge
423 establishment is to delineate field boundaries, one would expect a greater level of
424 conservation of hedge plants than within the field (Burel 1996; Louppe and Yossi 2000). Yet,
425 the decreases in tree density and basal area over the last two decades may reflect an
426 intensification of woody plant exploitation (Lericollais 1990). The trends observed in shrub
427 density confirm population stability, but again the slight downward trend in density that
428 occurred in the last decade and the progressive decrease in shrub basal area is consistent with
429 an increase in coppicing and exploitation pressure in spite of reported adoption by local
430 farmers of “Assisted Natural Regeneration” (RNA) management practices (Botoni et al.
431 2010). The density reduction of the two perennial herbaceous species, *A. gayanus* and *A.*
432 *sieberiana*, is also consistent with an increase in exploitation pressure, by cutting the haulms
433 of *A. gayanus* as construction material and grazing the dry season regrowth of both grasses.
434 Indeed, these declines in perennial density over the last decade is inconsistent with the trend
435 toward more frequent large rainfall events (Panthou et al. 2014) that should favor perennials
436 to the detriment of annual herbaceous species.

437

438 3.3.A denser web of field hedges resulting from land tenure fragmentation.

439 Among the 12 sample hedges observed in 2015, all were already in place in 1992, 8 in 1975
440 and 3 in 1950 (**Fig SI 2**), field hedges are thus persistent features of the landscape. Only one
441 of the hedges sampled in 1996 was dismantled by 2010 because of local soil erosion, and
442 among the 1006 km of hedges mapped in 1992 only 176 km were not observed in 2013-2016
443 (**Table 1**). Merging neighbor fields may eliminate some hedges but more frequent splitting of
444 fields generates new hedges and 1591 km of hedges were created between 1992 and 2013-
445 shares2016. Splitting a field in two should geometrically increase hedge length by a quarter
446 (in a square shaped field) to a third (in a rectangular field that is twice as long as wide). In 24
447 years, the overall length proportion of new versus old hedges reaches 1.58 which would
448 correspond to a bit less than two splits of each existing field on average, resulting in the
449 fragmentation of 1996 fields into 3 fields, on average. Indeed, when households divide, it is
450 more common for all fields managed by the family to be split equally among successors to
451 ensure that the quality of land is divided equally (Turner and Moumouni 2018a). Yet not all
452 field splits result in a woody field hedge, a tacit agreement on a demarcation line may suffice.
453 This appears to apply in the fields cropped by Fulani agro-pastoralists who do not have
454 primary rights on the land they crop (Sidikou 1997).

455

4563.4. **The increased contribution of field hedges to agricultural land woody plant population.**

457 The increased contribution of the hedges to the woody plants population at the landscape
458 scale mainly comes from the densification of the web of hedges with their total length
459 increasing by a factor of 2.4 in 24 years but is also influenced by the small changes in woody
460 population structure as assessed on the sample of twelve hedge monitored from 1996 to 2015.
461 The resulting contribution of hedges to woody plant density increased by 2.5, from 4.6 in
462 1995 to 9.2 plant ha⁻¹ in 2016, mostly due to increasing shrub numbers with crown cover only
463 increasing by a factor of 1.9 (**Table 3**). The hedge contribution to woody plant basal area
464 tripled, so that leaf mass multiplied by 4.1 and wood mass multiplied by 12.7. Whether these
465 increased contributions of field hedges to woody plant population compensate or not the
466 reduction of trees and shrubs within fields and fallows still needs to be assessed. And if
467 hedges do compensate for the losses within fields and fallows , would they further
468 compensate, at the landscape scale, the decimation of the woody plants of the patterned
469 vegetation of the plateau rangelands (Bouzou Moussa et al 2009; Abdourhamane Touré et al.
470 2010)?

471

472 **4. Conclusions**

473

474 In the monitoring of 12 sampled hedges from 1996 to 2015 no significant overall trends were
475 found for mean density, crown cover or basal area of the woody plants. Instead opposite
476 trends were observed between tree and shrubs, between dominant species and, above all,
477 among hedge sites. Selective cutting and coppicing would explain these locally specific
478 trajectories that are not explained by rainfall variations but likely from variable exploitation
479 pressure over the last decade. Indeed, the hedge's function as a land tenure marker that
480 ensures long life to most hedges does not prevent their exploitation through gathering (fruits,
481 dead wood, barks), pruning, and selective coppicing or cutting. The steady increase in rural
482 human population density is leading to a fragmentation of the agricultural lands through the
483 splitting of fields among successors at each generation. This fragmentation has led to an
484 expansion of woody field hedges by a factor 2.4 over the 24 years of the survey, equivalent to
485 a mean annual increase of 3.7%. The densification of the web of field hedges drives mean
486 annual increases in the contribution of field hedges to the woody plant population of the
487 agricultural land in the study area extending over 435 km². The field hedge woody plant
488 increased density by 3.9%, basal areas by 5.4% , crown cover by 2.7% , leading to 6.1%
489 annual increase in foliage mass and 8.8% increase in wood mass.

490

491 **Acknowledgements**

492 The field work was first carried out under an ILRI (International Livestock Research Institute,
493 Nairobi, Kenya) research project, more recent observations were funded by the AMMA-
494 CATCH observatory (<http://www.amma-catch.org>). The authors are very grateful to the rural
495 people of the Dantiandou district for their hospitality and for their patience with our endless
496 inquiries. They are also grateful to Laurent Kergoat and two anonymous reviewers for their
497 careful and inspirational review of the manuscript.

498

499 **References**

- 500 Abdourhamane Touré A., Guillon R., Garba Z., Rajot J.L., Petit C., Bichet V., Durand A.,
501 Sebag D., 2010. Evolution des paysages Sahéliens au cours des six dernières
502 décennies dans la région de Niamey : de la disparition de la brousse tigrée à
503 l'encroustement de surface des sols. *Pangea*, **47/48** : 35-40.
- 504 Achard F., Hiernaux P., Banoin M., 2001. Les jachères fourragères naturelle et améliorée en
505 Afrique de l'Ouest. In : *La jachère en Afrique tropicale : de la jachère naturelle à la*
506 *jachère améliorée. Le point des connaissances.* (Eds Floret C., Pontanier R.), John
507 Libbey Eurotext, Paris: 201-240.
- 508 Ayuk E.T., 1997. Adoption of agroforestry technology: The case of live hedges in the Central
509 Plateau of Burkina Faso. *Agricultural Systems* **54** (2): 189-206.
- 510 Ballouche A., Rasse M., 2007. L'homme, artisan des paysages de savane. *Pour la science*,
511 **358** :56-61.
- 512 Banzhaf J., Leihner D.E., Buerkert A. et Serafini P.G., 1992. Soil tillage and windbreak
513 effects on millet and cowpea: I. Wind speed, evaporation and wind erosion. *Agron. J.*,
514 **84**: 1056- 1060.
- 515 Boffa J.-M., 1995. Productivity and management of agroforestry parklands in the Sudan zone
516 of Burkina Faso, West Africa. PhD thesis Purdue University, West Lafayette, Indiana,
517 USA.
- 518 Boffa J.M., 1999. *Agroforestry Parkland Systems in Sub-Saharan Africa: FAO Conservation*
519 *Guide 34*, Roma.
- 520 Botoni E., Larwanou M., Reij C., 2010. La Régénération Naturelle Assistée (RNA) une
521 opportunité pour reverdir le Sahel et réduire la vulnérabilité des populations rurales.

522 In : Le projet majeur de Grande Muraille Verte de l’Afrique (Eds. Dia A., Duponnois
523 R.), IRD ed., Paris: 151-162.

524 Bouzou Moussa I., Descroix L., Maiga O.F., Gautier E., Adamou M.M., Esteves M., Souley
525 Yero K., Malam Abdou M., Mamadou I., Le Breton E., Abba B., 2011. Les
526 changements d’usage des sols et leurs conséquences hydrogéomorphologiques sur un
527 bassin versant endoréique sahélien. *Sécheresse*, **22**: 13-24. doi :
528 [10.1684/sec.2011.0297](https://doi.org/10.1684/sec.2011.0297)

529 Bouzou Moussa I., Maiga O.F., Ambouta J.M.K., B.Sarr, Descroix L., Adamou M.M., 2009.
530 Les conséquences géomorphologiques de l’occupation du sol et des changements
531 climatiques dans un bassin-versant rural sahélien. *Sécheresse*, **20** (1) : 145-152

532 Brandt M, Hiernaux P., Rasmussen K., Mbow C., Kergoat L., Tagesson T., Ibrahim Y. Z.,
533 Wélé A., Tucker J.C., Fensholt R., 2016. Assessing woody vegetation trends in
534 Sahelian drylands using MODIS based seasonal metrics. *Remote Sensing of Envir.*,
535 **183**: 215-222

536 Burel F., 1996. Hedgerows and their role in agri-cultural landscapes. *Critical Review in Plant*
537 *Sciences* **15**: 169–190.

538 Chamard P.C. et Courel M.F., 1999. La forêt sahélienne menacée. *Sécheresse*, **10**, 11-18.

539 Cissé M.I., 1980. Production fourragère de quelques arbres sahéliens : relation entre la
540 biomasse foliaire maximale et divers paramètres physiques. In : Les fourrages ligneux
541 en Afrique : état actuel des connaissances, (Ed. Le Houérou), Addis Abeba, ILCA: 203-
542 208

543 Depommier D., 1991. Propagation et comportement d'espèces à usages multiples en haies vives
544 pour la zone sahélo-soudanienne: résultats préliminaires d'essais menés à Gonse et
545 Dinderesso (Burkina Faso). In: *Physiologie des arbres et arbustes en zones semi-arides*.
546 (Eds. Riedacker, A., Dreyer E., Pafadnam C., Bory G.), Nancy, France : 155-165

547 Diack M., Sene M., Badiane A.N., Diatta M. , Dick R.P., 2000. Decomposition of a native
548 shrub (*Piliostigma reticulatum*) litter in soils of Semiarid Senegal. *J. of Arid Soil*
549 *Research and Rehabilitation* **14** (3): 205-218. DOI:[10.1080/089030600406626](https://doi.org/10.1080/089030600406626)

550 Diakhaté S., Gueye M., Chevallier T., Diallo N.H., Assigbetse K., Masse D., Sembène M.,
551 Ndour Y., Dick R.P., Chapuis-Lardy L., 2016. How the native woody shrub
552 *Piliostigma reticulatum* affects soil microbial communities when growing millet in
553 semi-arid Senegal. *J. of Arid Envir.* DOI :[10.1016/j.jaridenv.2016.01.010](https://doi.org/10.1016/j.jaridenv.2016.01.010)

554 Dick R., Sene M., Diack M., Khouma M., Badiane A., Samba S. A., Diedhiou I, Lufafa A.,
555 Dossa E., Kizito F., Diedhiou S., Noller J., Dragila M., 2010. The native

556 shrubs *Piliostigma reticulatum* and *Guiera senegalensis*: The unrecognized potential
557 to remediate degraded soils and optimize productivity of Sahelian agroecosystems. In :
558 *Le projet majeur africain de la Grande Muraille Verte : Concepts et mise en*
559 *œuvre.*(Eds : Dia, A., Duponnois, R.), IRD Éditions.
560 doi :10.4000/books.irdeditions.2126

561 d'Herbès J.M., Valentin C., Thiéry J.M., 1997. La brousse tigrée au Niger: synthèse des
562 connaissances acquises. Hypothèses sur la genèse et les facteurs déterminant les
563 différentes structures contractées. In : *Fonctionnement et gestion des écosystèmes*
564 *forestiers contractés sahéliens.*(Eds. d'Herbès J.M., Ambouta J.M.K., Peltier R.), John
565 Libbey Eurotext, Paris : 131-152.

566 Favreau G., Leduc C., Marlin C., Dray M., Taupin J.-D., Massault M., Le Gal La Salle C.,
567 Babic M., 2002. Estimate of recharge of a rising water table in semiarid Niger from ³H
568 and ¹⁴C Modeling. *Ground water*, **40** (2): 144-151

569 Hiernaux P., 1980. L'inventaire du potentiel fourrager des arbres et des arbustes d'une région
570 du Sahel malien. Méthodes et premiers résultats. In : *Les fourrages ligneux en*
571 *Afrique: état actuel des connaissances*, (Ed. Le Houérou), Addis Abeba, ILCA: 1995-
572 201

573 Gijssbers H.J.M., Kessler J.J., Knevel M.K., 1994. Dynamics and natural regeneration of
574 woody species in farmed parklands in the Sahel region (Province of Passoré, Burkina
575 Faso). *Forest Ecology and Management*, **64**: 1-12.

576 Henry M., Picard N., Trotta C., Manlay R.J., Valentini R., Bernoux M., Saint-André L., 2011.
577 Estimating tree biomass of sub-saharan African forests: a review of available
578 allometric equations. *Silva Fennica*, **45**, 3B: 477-569

579 Hiernaux P., Ayantunde A.A., 2004. The Fakara: a semi-arid agro-ecosystem under stress.
580 Report of research activities of the DMP-GEF programme (GEF/2711-02-4516), ILRI,
581 Nairobi, Kenya, 95p.

582 Hiernaux P., Diarra L., Trichon V., Mougin E., Soumaguel N., Baup F., 2009. Woody plant
583 population dynamics in response to climate changes from 1984 to 2006 in Sahel
584 (Gourma, Mali). *Journal of Hydrology*, **375** : 103–113.

585 Hiernaux P., Diawara M.O., 2014. Livestock: recyclers that promote the sustainability of
586 smallholder farms. *Rural 21*, **4**: 9-1.

587 Hiernaux P., Le Houérou H.N., 2006. Les parcours du Sahel. *Sécheresse*, **17** (1-2) : 1-21, 51-
588 71.

589 Hiernaux P., Turner M. D., 2002. The influence of farmer and pastoralist management
590 practices on desertification processes in the Sahel. In: Global Desertification. Do
591 Humans cause deserts? (Eds. Reynolds J.F., Smith, M.S.), Dahlem Univ. Press,
592 Berlin: 135–148.

593 Idrissa S., Habou R., Issaharou I. M., Ali M., Saadou M., 2017. Biodiversity and Structure of
594 Woody Plants of Sahelian Rangelands of Baban Rafi, Niger. *Intern. J. of Biology*. 9. 1.
595 10.5539/ijb.v9n4p1.

596 INS-Niger, 2013. Présentation des résultats préliminaires du quatrième (4^{ième}) recensement
597 général de la population et de l’habitat (RGP/H) 2012. Min. Finances, Institut National
598 de la Statistique, Niamey, Niger, 10pp

599 Lavigne-Delville Ph., 2003. When farmers use “pieces of paper” to record their land
600 transactions in francophone rural Africa : Insights into the dynamics of institutional
601 innovation. In: Securing Land Rights in Africa (Eds. Benjaminsen T.A., Lund C.),
602 London/Bonn, Franck Cass/EADI: 89-108.

603 Lebel T., Ali A., 2009. Recent trends in the
604 Central and Western Sahel rainfall regime (1990–2007). *Journal of Hydrology*, 375, 1-
605 2 : 52–64.

606 Lericollais A., 1990. La gestion du paysage, Sahélisation, surexploitation et délaisement des
607 terroirs sereer au Sénégal. In : La dégradation des paysages en Afrique de l’Ouest. (ed.
608 Richard J.F.) :151-169.

609 Levasseur V., 2003. L’utilisation des haies vives améliorées dans le cercle de Ségou, au Mali
610 : Le signe d’une société en mutation. Thèse de Ph.D. en Agronomie Tropicale,
611 Université Laval, Québec. 241 p.

612 Louppe D., Yossi H., 2000. Les haies vives et défensives en zones sèche et sub-humide
613 d’Afrique de l’Ouest. In : La jachère en Afrique tropicale. De la jachère naturelle à la
614 jachère améliorée, le point des connaissances(Eds. Floret C.,Pontanier R.), John
615 Libbey Eurotext, Paris: 293-309

616 Mahamane A., Saadou M., Baïna Danjimo M., Karim S., Bakasso Y., Diouf A., Arzika, T.,
617 2009. Plants diversity in Niger: state of the present studies. *Ann. Univ. of Lomé*,
618 Togo : 81-93.

619 Michels S.K., Lamers J.K.A., Buerkert A., 1998. Effects of windbreak species and mulching
620 on wind erosion and millet yield in the Sahel. *Expl. Agric.* : 229-236.

621 Nouvellet Y., Kassambara A, Besse F., 2006. Le parc à Karités au Mali : inventaire, volume,
622 houppier et production fruitière. *Bois et forêt des tropiques*, **287**: 5-20.

622 Osbahr H., 2001. Livelihood strategies and soil fertility at Fandou Béri, southwestern Niger.
623 PhD thesis Geography, Univ. College London, UK, 414p.

624 Panthou G., Vischel T., Lebel T., 2014. Recent trends in the regime of extreme rainfall in
625 central Sahel Int. J. Climatol., **34**: 3998-4006

626 Péliissier P., 1953. Les paysans Sérères. Essai sur la formation d'un terroir du Sénégal. *Les*
627 *Cahiers d'Outre-Mer*, **6** (22): 105-127.

628 Péliissier P., 1995. Campagne africaine en devenir. Arguments, Paris. 318 p.

629 Peltier R.J., Dessard H., Gado Alzouma R., Ichaou A., 2009. Bilan après quinze ans de
630 gestion communautaire d'une forêt villageoise de l'Ouest nigérien. *Sécheresse*, **20**
631 (4) : 383-387

632 Picard N., Bar-Hen A., 2007. Estimation of the density of a clustered point pattern using a
633 distance method. *Envir. and Ecol. Stat.*, **13** (3): 341-353.

634 Portères R., 1965. Le Caractère magique originel des haies vives et de leurs constituants.
635 *Journal d'agriculture tropicale et de botanique appliquée*, **12** (6-8): 253-291.

636 Poux X., Narcy J.B., Romain B., 2009. Réinvestir le saltus dans la pensée agronomique
637 moderne: vers un nouveau front éco-politique? *L'Espace Politique*, 9 (3). DOI :
638 10.4000/espacepolitique.1495

639 Raynaut C., 2001. Societies and nature in the Sahel: ecological diversity and social dynamics.
640 *Global Envir. Change*, **11** : 9-18.

641 San Emeterio J.L., Alexandre F., Andrieu J., Génin A., Mering C., 2013. Changements socio-
642 environnementaux et dynamiques des paysages ruraux le long du gradient
643 bioclimatique nord-sud dans le sud-ouest du Niger (régions de Tillabery et de
644 Dosso). *Vertigo*, **13** (3), DOI : 10.4000/vertigo.14456

645 Sanogo D., Dia Y.K., Ayuk E., Pontanier R., 2000. Adoption de la haie vive dans le bassin
646 arachidier du Sénégal. In : La jachère en Afrique Tropicale : rôles, aménagements,
647 alternatives (Eds. Floret C., Pontanier R.), Vol 1., John Libbey Eurotext, Paris :733-
648 740.

649 Seignobos C., 1980. Des fortifications végétales dans la zone soudano sahéliennes (Tchad et
650 Nord Cameroun). Cah. ORSTOM, série Sc. Hum., **12** (3-4): 191-222.

651 Sidikou H.A., 1997. Droits d'usages traditionnels locaux et demande externe des populations
652 urbaines au Niger. In : Fonctionnement et gestion des écosystèmes forestiers
653 contractés sahéliens (Eds. d'Herbès J.M., Ambouta J.M.K., Peltier R.), J. Libbey
654 Eurotext, Paris: 3-14.

655 Smolikowski B., Puig H., Roose E., 2001. Influence of soil protection techniques on runoff,
656 erosion and plant production on semi-arid hillsides of Cabo Verde. *Agric. Ecos. and*
657 *Envir.*, **87**: 67-80

658 Tong X., Brandt M., Hiernaux P., Herrmann S.M., Tian F., Prishchepov A.V., Fensholt R.,
659 2017. Revisiting the coupling between NDVI trends and cropland changes in the Sahel
660 drylands: la case study in western Niger. *Remote Sensing of Envir.*, **191**: 286–296

661 Trichon V., Hiernaux P., Walcker R., Mougin E., 2018. The persistent decline of patterned
662 woody vegetation: The tiger bush in the context of the regional Sahel greening trend.
663 *Glob Change Biol.* 00:1–16. <https://doi.org/10.1111/gcb.14059>

664 Turner M.D., McPeak J., Ayantunde A.A., 2014. The role of livestock mobility in the
665 livelihood strategies of rural peoples in semi-arid West Africa. *Human Ecol.* **42** (2):
666 231–247

667 Turner M.D., Hiernaux P., 2002. The use of herders' accounts to map livestock activities
668 across agropastoral landscapes in Semi-Arid Africa. *Landscape Ecology*, **17**: 367 –
669 385.

670 Turner M. D., Hiernaux P., 2015. The effects of management history and landscape position
671 on inter-field variation in soil fertility and millet yields in southwestern Niger. *Agric.*
672 *Ecos. and Envir.*, **211**: 73-83

673 Turner M.D., Hiernaux P., Schlecht E., 2005. The Distribution of grazing pressure in relation
674 to vegetation resources in semiarid West Africa: The Role of Herding. *Ecos.* **8** (6):
675 668–681.

676 Turner M.D., Moumouni O., 2018a. Mosaics of property : control of village land in West
677 Africa. *The Journal of Peasant Studies*.
678 [https://doi.org/10.1080:03066150.2018.1439931](https://doi.org/10.1080/03066150.2018.1439931)

679 Turner M.D., Moumouni O., 2018b. The dividing of fields in Sudano-Sahelian West Africa:
680 the roles of soil fertility variation and legal doctrine. *Land Use Policy*, **77** : 362-374

681 Valet S., 2000. Nouvelle stratégie d'éco-développement durable par la gestion et la
682 valorisation du report hydrique. *Sécheresse*, 4 (11) :239-247.

683 Yossi H., Kaya B., Traoré C.O., Niang A., Butaré I., Levasseur V., Sanogo D., 2006. Les
684 haies vives au Sahel. Etat des connaissances et recommandations pour la recherche et
685 le développement. ICRAF, Nairobi, Kenya, Occ. Paper 6, 52p.